

¹ A Blueprint for Process-Based Modeling of ² Uncertain Hydrological Systems

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3 **Abstract.** We present a probability based theoretical scheme for build-
4 ing process-based models of uncertain hydrological systems, thereby unify-
5 ing hydrological modeling and uncertainty assessment. Uncertainty for the
6 model output is assessed by estimating the related probability distribution
7 via simulation, thus shifting from one to many applications of the selected
8 hydrological model. Each simulation is performed after stochastically per-
9 turbing input data, parameters and model output, this latter by adding ran-
10 dom outcomes from the population of the model error, whose probability dis-
11 tribution is conditioned on input data and model parameters. Within this
12 view randomness, and therefore uncertainty, is treated as an inherent prop-
13 erty of hydrological systems. We discuss the related assumptions as well as
14 the open research questions. The theoretical framework is illustrated by pre-
15 senting real-world and synthetic applications. The relevant contribution of
16 this study is related to proposing a statistically consistent simulation frame-
17 work for uncertainty estimation which does not require model likelihood com-
18 putation and simplification of the model structure. The results show that un-
19 certainty is satisfactorily estimated although the impact of the assumptions
20 could be significant in conditions of data scarcity.

1. Introduction

21 Process-based modeling has been a major focus for hydrologists for four decades al-
22 ready. In fact, more than forty years passed since *Freeze and Harlan* [1969] proposed
23 their “physically-based digitally simulated hydrologic response model”, which set the ba-
24 sis for detailed process-based simulation in hydrology. The terms “physically-based” and
25 “process-based” models are often used interchangeably, in contrast to purely empirical
26 models. Other times, “process-based” is regarded to include a family of models broader
27 than “physically-based”. In fact, through the years it has become clear that there are
28 no purely “physically-based” models for large hydrological systems. All models include
29 assumptions and simplifications that depart from pure deductive physics and thus the
30 term “process-based” is more accurate and general.

31 A comprehensive review of the research activity related to physically-based models was
32 presented by *Beven* [2002]. Perhaps one of the most known process-based models in hy-
33 drology is the spatially-distributed *Système Hydrologique Européen* (SHE, see *Abbot et al.*
34 [1986]), which has been the subject of many contributions. Another relevant contribution
35 was given by *Reggiani et al.* [1998, 1999] who introduced the concept of “representative
36 elementary watershed”. Many other approaches were recently proposed which refer to
37 process-based models in general. In fact, in the past ten years, process-based modeling,
38 in contrast to empirical modeling, has been one of the targets of the well known “Pre-
39 diction in Ungauged Basins” (PUB; see *Kundzewicz* [2007]) initiative of the International
40 Association of Hydrological Sciences (IAHS).

41 Process-based models are almost always formulated in deterministic form, by setting
42 up a set of mathematical equations. However, during the last four decades it became
43 increasingly clear that deterministic models in hydrology are never accurate and imply
44 uncertainty whose estimation is important for real world decision making (see, for in-
45 stance, *Grayson et al.* [1992]; *Beven* [1989, 2001]). Some authors expressed their belief
46 that uncertainty in hydrology is epistemic and therefore can be in principle eliminated
47 through a more accurate representation of the related processes [*Sivapalan et al.*, 2003].
48 However, recent research suggested that uncertainty is unavoidable in hydrology, origi-
49 nating from natural variability and related to inherent unpredictability in deterministic
50 terms, which is typically referred to as randomness (see, for instance, *Montanari et al.*
51 [2009]; *Koutsoyiannis et al.* [2009]). The latter concept implies that to produce a fully
52 deterministic model that would eliminate uncertainty is impossible and that modeling
53 schemes need to explicitly recognise its role [*Beven*, 2002].

54 Indeed, recent contributions were proposed where deterministic hydrological modeling
55 was efficiently coupled with uncertainty assessment. The most relevant example is the
56 Generalised Likelihood Uncertainty Estimation (GLUE) [*Beven and Binley*, 1992], where
57 multiple modeling schemes are retained provided they are behavioral in the face of uncer-
58 tainty. GLUE was long discussed and sometimes criticized for using informal approaches
59 for statistical inference and in particular for computing the model likelihood (see, for in-
60 stance, *Stedinger et al.* [2008] and *Mantovan and Todini* [2006]), but was used in several
61 practical applications. Relevant recent contributions to GLUE were given by *Liu et al.*
62 [2009], who proposed the limits of acceptability approach to retain behavioral simulations
63 (see also *Winsemius et al.* [2009]), and *Krueger et al.* [2010], who explicitly considered the

64 contribution of data, parameter and model uncertainty via ensemble simulation. How-
65 ever, GLUE still suffers from subjectivity, mainly related to the identification of behavioral
66 models and probability estimation for their output [*Stedinger et al.*, 2008], which in GLUE
67 is obtained through (possibly informal) likelihood estimation for any candidate model.

68 A second relevant approach to uncertainty assessment was proposed by *Krzysztofowicz*
69 [2002] who introduced the Bayesian Forecasting System (BFS), which aims to estimate the
70 probability distribution for a future river stage or river flow. The method is based on the
71 preliminary identification of a prior distribution for the unknown future variable, which is
72 obtained by approximating the river flow process with a linear model, therefore expressing
73 future observations depending on past ones. Then, the posterior distribution is obtained
74 by including the information provided by the hydrological model prediction. The proba-
75 bility distributions are estimated in the Gaussian domain by normalizing predictors and
76 predictand through the normal quantile transform [*Krzysztofowicz*, 2002]. The method
77 assumes that the dominant source of uncertainty is related to rainfall forecasting, thus
78 focusing on a very specific type of application. The above assumption implies that hydro-
79 logical uncertainty is estimated by introducing restrictive approximations. In particular,
80 parameter uncertainty for the hydrological model is neglected.

81 Estimation of hydrological uncertainty is also the target of the BATEA method [*Kavetski*
82 *et al.*, 2006], where a novel approach is introduced to account for all sources of data
83 uncertainty. In particular, rainfall uncertainty is accounted for by introducing a rainfall
84 multiplier. The probability distribution for model parameters is estimated through the
85 Bayes theorem and therefore a formulation for the model likelihood needs to be identified.

86 For example, *Kavetski et al.* [2006] adopted a likelihood function depending on the sum
87 of squared residuals.

88 In recent times, there has been a renewed interest for multi-model approaches, which
89 estimate unknown hydrological variables by averaging outputs from several models. This
90 possibility is also offered by GLUE [*Krueger et al.*, 2010]. Another relevant example is
91 the Maximum Likelihood Bayesian Model Averaging (MLBMA, see *Neuman* [2003]; *Ye*
92 *et al.* [2008]). Multi-model techniques may require likelihood estimation to derive the
93 probability that each model is correct.

94 The above considerations show that model likelihood estimation is a key step for many
95 uncertainty assessment methods. It is well known that likelihood computation for hy-
96 drological models is a very challenging task, due to the complex structure of the model
97 error which makes its statistical description complicated. Interesting contributions were
98 recently proposed by *Schoups and Vrugt* [2010] and *Pianosi and Raso* [2012] who pro-
99 posed innovative likelihood formulations. However, they are still based on assumptions
100 that may be restrictive in some practical applications, like the use of a model bias correc-
101 tion factor in *Schoups and Vrugt* [2010] and the hypothesis of independence for the model
102 error in *Pianosi and Raso* [2012]. Moreover, there is a drawback related to the use of
103 the likelihood to estimate the reliability of a model in hydrology: in fact, the likelihood is
104 usually estimated in calibration, as it is done in statistics, but is used to assess uncertainty
105 of out-of-sample predictions. Therefore it is implicitly assumed that model performances
106 in calibration are analogous to those experienced in the evaluation period. Namely, one
107 assumes that the model errors during calibration are statistically representatives of those
108 that will be experienced in applications. Actually, this assumption is valid only under the

109 condition that the hydrological model is stationary and not overparametrised, but fails in
110 all instances in which the actual model reliability is expected to deteriorate with respect
111 to calibration (as it frequently happens in hydrology). This limitation, in the context
112 of GLUE, is recognised by *Beven* [2006], p. 27. To overcome it, likelihood should be
113 computed by running the model, after optimizing its parameters, during an evaluation
114 period. Namely, one should refer to data that were not used in calibration and similar
115 conditions with respect to those that are expected in applications.

116 An approach to hydrological uncertainty assessment which does not require likelihood
117 estimation was presented by *Götzinger and Bárdossy* [2008] who assumed that the model
118 error is given by the sum of the random components due to input uncertainty and pro-
119 cess description uncertainty. To estimate this latter contribution, they assumed that the
120 standard deviation of the random contribution of a certain process (model structural
121 uncertainty) to the total uncertainty is proportional to the sensitivity of the output to
122 the related parameter group. The above assumptions may not be satisfied in practical
123 applications (see, for instance, *Beven* [2006]).

124 Likelihood computation might also be avoided by using data assimilation methods, for
125 which a comprehensive review, from a system-perspective, was presented by *Liu and Gupta*
126 [2007]. In fact, the Bayesian uncertainty assessment method developed by *Bulygina and*
127 *Gupta* [2009, 2010, 2011] assumes that the hydrological system evolves in time according
128 to a first-order Markov state-space process and the relationships among the relevant vari-
129 ables (inputs, states and outputs) is represented through the direct estimation of their
130 joint probability density function. This latter takes uncertainty into account and is con-
131 ditioned on both the observed data and the available conceptual understanding of system

132 physics, therefore obtaining a flexible and statistically consistent approach. *Bulygina and*
133 *Gupta* [2010] note that additional research is needed to make the method applicable to
134 complex system models. Moreover, if no or weakly informative prior is used, any predic-
135 tion is mainly based on the conditions observed during the considered observation period
136 only, and therefore particular care should be used when extrapolating to out-of-sample
137 conditions.

138 The purpose of this paper is to introduce a novel methodological scheme for estimating
139 the probability distribution of the output from a process-based (deterministic) hydrolog-
140 ical model. The distinguishing feature of the approach herein proposed is that likelihood
141 computation can be avoided, without imposing any restriction to model complexity, there-
142 fore complementing the features of the techniques reviewed above. Conversely, the most
143 significant limitation is that the probability distributions of input data, model parameters
144 and model error are needed as input information. It is well known that their definition
145 is still a challenging task in practical applications. The underlying theory is derived in
146 a probabilistic framework, in which Bayesian concepts can be introduced to take into
147 account prior information. Statistical consistency of the scheme is ensured by introducing
148 assumptions whose reliability is discussed below. The scheme itself is based on converting
149 a deterministic hydrological model into a stochastic one, therefore incorporating random-
150 ness in hydrological modeling as a fundamental component. In fact, in our framework
151 uncertainty is recognized as an inherent property of the water cycle, taking into account
152 randomness of atmospheric processes, drop paths, soil properties, turbulence in fluid me-
153 chanics and many others.

154 In the next Section of the paper we provide more details on the rationale for stochastic
155 process-based modeling. The third section of the paper is dedicated to the theory under-
156 lying the new blueprint that we are proposing. The fourth section describes the practical
157 application. The fifth section reviews the underlying assumptions and their limitations,
158 while in the sixth section two examples of application are presented. In the seventh section
159 we discuss the value of uncertainty estimation as a learning process. Finally we discuss
160 open research questions and draw some conclusions.

2. The rationale for stochastic process-based modeling

161 In a deterministic model the outcomes are precisely determined through known rela-
162 tionships among states and events, without any room for random variation. A given
163 input (including initial and boundary conditions) will always produce the same output
164 and therefore uncertainty is not taken into account in a formal manner. Uncertainty as-
165 sessment, when needed, is often carried out indirectly, e.g., by post processing the results.
166 Such separation of targets has been favoured by the illusion that uncertainty can be elim-
167 inated by refining deterministic modeling (see, for instance, *Sivapalan et al.* [2003]). Such
168 refining has commonly been envisaged through a “reductionist” approach, in which all
169 heterogeneous details of a catchment would be modeled explicitly and the modeling of de-
170 tails would provide the behavior of the entire system (for an extended discussion see *Beven*
171 [2002]). However, some researchers have pointed out that this is a hopeless task [*Savenije,*
172 2009]. Indeed, physical processes governing the water cycle involve inherent randomness;
173 for instance, meteorological processes are governed by the laws of thermodynamics, which
174 are, essentially, statistical physical laws. Moreover, some degree of approximation is un-
175 avoidable in process-based modeling in hydrology [*Beven, 1989*]. In fact, physical laws give

176 simple and meaningful descriptions of problems in simple systems, but their application
177 in hydrological systems demands simplification, lumping and statistical parameterization
178 *Beven* [1989], and sometimes even replacing by conceptual or statistical laws (e.g. the
179 Manning formula). Therefore, uncertainty in hydrology is not just related to temporary
180 knowledge limitations (epistemic uncertainty) but it rather reflects inherent randomness
181 and therefore it is unavoidable (see *Koutsoyiannis et al.* [2009]). Thus, the traditional
182 deterministic form of process-based modeling in hydrology is a relevant limitation per se
183 which should be overcome by incorporating uncertainty modeling in a fully integrated
184 approach.

185 In fact, we believe that recognizing uncertainty as an essential attribute of the water
186 cycle, which needs to be respected in process-based models, is not just a nuisance. In our
187 view, uncertainty estimation is not the remedy against limited representativity of deter-
188 ministic schemes (which some may believe to be a transient weakness of current models
189 that would be cleared up in the future), but rather a way to fully take into account and
190 reproduce in a process-based framework the dynamics of hydrological systems. There are
191 many possible alternatives to deal with uncertainty thereby overcoming the limitations of
192 purely deterministic approaches, including subjective methods like fuzzy logic, possibil-
193 ity theory, and others [*Beven*, 2009; *Montanari*, 2007]. We believe that one of the most
194 comprehensive, elegant and complete ways of dealing with uncertainty is provided by the
195 theory of probability. In fact, probabilistic descriptions allow predictability (supported
196 by deterministic laws) and unpredictability (given by randomness) to coexist in a unified
197 theoretical framework, therefore giving us the means to efficiently exploit and improve
198 the available physical understanding of uncertain systems [*Koutsoyiannis et al.*, 2009].

199 The theory of stochastic processes also allows the incorporation into our descriptions of
200 (possibly human induced) changes affecting hydrological processes [*Koutsoyiannis*, 2011],
201 by modifying their physical representation and/or their statistical properties (see, for in-
202 stance, *Merz and Blöschl* [2008a, b]). Finally, subjectivity and expert knowledge can
203 be taken into account in prior distribution functions through Bayesian theory [*Box and*
204 *Tiao*, 1973]. Therefore, in our opinion, a theoretical setting needs to be established where
205 probability-based modeling of uncertainty is an essential piece of possibly complex deter-
206 ministic models. Such a setting should be flexible enough to allow the deterministic model
207 to increase in complexity therefore reducing epistemic uncertainty as much as possible in
208 the future, while retaining the essential role of inherent randomness.

209 In view of the above considerations and in agreement with *Beven* [2002], we believe
210 that a new blueprint should be established, which should be built on a key concept that is
211 actually well known: it is stochastic process-based modeling, which needs to be brought
212 to a new light in hydrology. Here the term “stochastic” is used to collectively represent
213 probability, statistics and stochastic processes. We formalize the theoretical framework
214 for the application of this type of approach here below.

3. Formulating a Process-Based Model Within a Stochastic Framework

215 In this section we show how a deterministic model can be converted into an essential part
216 of a wider stochastic approach through an analytical transformation. Such a conversion
217 is necessary to understand how the probability distribution of the model output can be
218 estimated by simulation, without necessarily requiring likelihood computation. From an
219 analytical point of view, while the deterministic formulation of the model transforms the
220 values of the inputs into an output value, the stochastic version of the model acts on

221 probability densities, rather than single values, of the inputs, producing a probability
222 density of the output. That is, the deterministic model acts on values of variables while
223 the stochastic model acts on probability densities thereof. From a numerical point of view,
224 a deviation (error term) from a single-valued relationship is introduced and the density
225 of the output is calculated by repeated applications of the single-valued (deterministic)
226 version of the model, where the model output is stochastically perturbed to account for
227 uncertainties in a statistically consistent framework. The scheme presented here focuses
228 on the conversion of a single deterministic model into a stochastic model. However, a
229 multi-model extension, which uses more than one deterministic model, is straightforward
230 and will be discussed below.

231 In what follows, we use the following definitions:

232 • Input uncertainty: it is defined as the uncertainty in the data input to the model,
233 which is quantified by an underlying probability distribution. It is related to observation
234 methods and networks.

235 • Parameter uncertainty: it is defined as the uncertainty in the model parameter vec-
236 tor. It is mainly related to model structure, calibration method and consistency of the
237 underlying data base.

238 • Model error: for a given model, input data and parameter vector, it is defined as
239 the difference between model simulation and the corresponding data value. It is mainly
240 related to model inability to reproduce the related real processes (model structural error).
241 Here, it is assumed to resemble all the uncertainties that are not included in input data
242 and parameter uncertainty.

243 • Prediction uncertainty: for a given model (or set of models in a multi-model frame-
 244 work), it is defined as the uncertainty in the prediction of the true value of a given
 245 hydrological variable. It is quantified by the probability distribution of the variable to be
 246 predicted and is typically expressed also in the form of prediction limits of the simulation.
 247 These latter quantify the range for the variable within which the true value falls with
 248 probability equal to the confidence level. Prediction uncertainty is formed up by input
 249 and data uncertainty and model error.

250 The analytical procedure to convert a deterministic model into a stochastic framework
 251 is rather technical and is expressed by equations (1) to (6) below. We would like to
 252 introduce the new blueprint with a fully comprehensible treatment for those who are not
 253 acquainted with (or do not like) stochastics. Therefore, the presentation is structured to
 254 allow the reader who is interested in the application only to directly jump to equations
 255 (7) and (8) without any loss of practical meaning.

256 Hydrological models are typically expressed through a deterministic formulation,
 257 namely, a single valued transformation. In general, it can be written as

$$Q = S(\Theta, X) \quad (1)$$

259 where Q is the model prediction which, in a deterministic framework, is implicitly assumed
 260 to equal the true value of the variable to be predicted. The mathematical relationship S
 261 represents the model structure, X indicates the input data vector and Θ the parameter
 262 vector. In the stochastic framework, the hydrological model is expressed in stochastic
 263 terms, namely [*Koutsoyiannis, 2010*],

$$f_Q(Q) = K_S f_{\Theta, X}(\Theta, X) \quad (2)$$

265 where f indicates a probability density function, and K_S is a transfer operator that is
 266 related to, and generalizes in a stochastic context, the deterministic model S . Within this
 267 context, Q indicates the true variable to be predicted, which is unknown at the prediction
 268 time and therefore is treated as a random variable. K_S can be generalized to represent a
 269 so-called stochastic operator, which implements a shift from one to many transformations
 270 S .

271 Note that, by starting from eq. (1) and (2) above, we assume that Q depends on input
 272 data X and parameters Θ through the model S . Therefore, f_Q (and thus uncertainty of Q)
 273 depends on $f_{X,\Theta}$ (and thus uncertainty of X and Θ) and model S uncertainty (through the
 274 operator K_S). It follows that the model error is assumed to resemble all the uncertainties
 275 that are not included in input data and parameter uncertainty, as it was noted in the
 276 definitions above. In principle, other uncertainty sources could be considered explicitly.
 277 For instance, dependence on the initial conditions and therefore their uncertainty can be
 278 easily included in eq.(1) and (2). In what follows it is omitted to simplify notation (note
 279 that initial conditions can be included in the input data vector X).

280 A stochastic operator can be defined by using a stochastic kernel $k(e, \Theta, X)$, with e
 281 reflecting a deviation from a single valued transformation. Here we will assume that e is a
 282 stochastic process, with marginal probability density $f_e(e)$, representing the model error
 283 according to the additive relationship

$$284 \quad Q = S(\Theta, X) + e. \quad (3)$$

285 Note that alternative structures for the model error can be defined, for instance by in-
 286 troducing multiplicative terms. The term e accounts for the model uncertainty that was
 287 discussed above.

288 The stochastic kernel introduced above must satisfy the following conditions:

$$289 \quad k(e, \Theta, X) \geq 0 \text{ and } \int_e k(e, \Theta, X) de = 1, \quad (4)$$

290 which are met if $k(e, \Theta, X)$ is a probability density function with respect to e .

291 Specifically, the operator K_S applying on $f_{\Theta, X}(\Theta, X)$ is then defined as [Lasota and
292 Mackey, 1985, p. 101]

$$293 \quad K_S f_{\Theta, X}(\Theta, X) = \int_{\Theta} \int_X k(e, \Theta, X) f_{\Theta, X}(\Theta, X) d\Theta dX. \quad (5)$$

294 Under the assumption that parameter uncertainty is independent of data uncertainty,
295 for the purpose of estimating the probability density $f_Q(Q)$ the joint probability distri-
296 bution $f_{\Theta, X}(\Theta, X)$ can be substituted by the product of the two marginal distributions
297 $f_{\Theta}(\Theta) f_X(X)$. Note that we are not excluding dependence among the single elements of
298 the input data as well as parameter vector, and also note that this assumption can be
299 removed and therefore does not affect the generality of the approach, as we discuss in
300 Section 5.

301 In view of this latter result, by combining eq. (2) and eq. (5), in which the model error
302 can be written as $e = Q - S(\Theta, X)$ according to eq. (3), we obtain

$$303 \quad f_Q(Q) = \int_{\Theta} \int_X k(Q - S(\Theta, X), \Theta, X) f_{\Theta}(\Theta) f_X(X) d\Theta dX. \quad (6)$$

304 At this stage we need to identify a suitable expression for $k(Q - S(\Theta, X), \Theta, X)$. Upon
305 substituting eq. (3) in eq. (6) and remembering that k is a probability density function
306 with respect to the model error e , we recognize that the kernel is none other than the
307 conditional density function of e for the given Θ and X , i.e., $f_e(Q - S(\Theta, X) | \Theta, X)$.

308 To summarise the whole set of analytical derivations expressed by equations (1) to (6)
309 we may see that we passed from the deterministic formulation of the hydrological model

310 expressed by eq. (1), i.e. (to replicate it for clarity),

$$311 \quad Q = S(\Theta, X) \quad (7)$$

312 to the stochastic formulation expressed by

$$313 \quad f_Q(Q) = \int_{\Theta} \int_X f_e(Q - S(\Theta, X) | \Theta, X) f_{\Theta}(\Theta) f_X(X) d\Theta dX \quad (8)$$

314 with the following meaning of the symbols:

315 - $f_Q(Q)$: probability density function of the true value of the hydrological variable to be
316 predicted;

317 - $S(\Theta, X)$: deterministic part of the hydrological model;

318 - $f_e(Q - S(\Theta, X) | \Theta, X)$: conditional probability density function of the model error. Ac-
319 cording to eq. (2) it can also be written as $f_e(e | \Theta, X)$;

320 - Θ : model parameter vector;

321 - $f_{\Theta}(\Theta)$: probability density function of model parameter vector;

322 - X : input data;

323 - $f_X(X)$: probability density function of input data.

324 In eq. (8) the conditional probability distribution of the model error $f_e(Q - S(\Theta, X) | \Theta, X)$
325 is conditioned on input data X and parameter vector Θ . Such formulation would be use-
326 ful if we needed to account for changes in time of the conditional statistics of the model
327 error (like, for instance, those originated by heteroscedasticity). On the other hand, if we
328 assumed that the model error is independent of X and Θ , then eq. (8) can be written in
329 the simplified form

$$330 \quad f_Q(Q) = \int_{\Theta} \int_X f_e(Q - S(\Theta, X)) f_{\Theta}(\Theta) f_X(X) d\Theta dX. \quad (9)$$

331 The presence of a double integral in eq. (8) and eq. (9) may induce the feeling to the
 332 reader that the practical application of the proposed framework is cumbersome. Actually,
 333 the double integral can be easily computed through numerical integration, namely, by
 334 applying a Monte Carlo simulation procedure that is well known and already used in
 335 hydrology (see *Koutsoyiannis* [2010]). The only problem is related to the computational
 336 requirement, which might become significant when dealing with complex models and large
 337 basins. We explain the numerical integration in the next section of the paper.

338 The above theoretical scheme is very general yet its formalism has been given in terms
 339 of converting a single deterministic model into a stochastic model. However, the gener-
 340 ality and flexibility of the approach allow for an extension to a multi-model framework.
 341 Multi-modeling schemes allow to test multiple working hypotheses and model structures
 342 thereby testing individual components of process-based models (for an extensive discus-
 343 sion see *Clark et al.* [2011]; for an example of application see *Krueger et al.* [2010]). A
 344 preliminary estimation of the weight w_i of each i -th model is necessary, which quanti-
 345 fies the importance of each model in the simulation process. The weight is related to
 346 model performances with respect to other candidate models (*Neuman* [2003], page 297,
 347 defines the weight as the probability that the model is correct), and can be estimated
 348 by using prior judgemental information. Uniform probability across different models is a
 349 reasonable working hypothesis, which should however be supported by expert knowledge
 350 to avoid that the same importance is given to models with much dissimilar predicting
 351 capabilities. The multi-model probability density function $f_Q(Q)$ can be written as

$$352 \quad f_Q(Q) = \sum_{i=1}^M w_i f_Q^{(i)}(Q) , \quad (10)$$

353 where M is the number of the considered models and $f_Q^{(i)}(Q)$ is the probability distribution
354 derived through eq. (8) or (9) for each single model i , $1 < i \leq M$.

355 It can be seen that methodological scheme introduced above does not require compu-
356 tation of the model likelihood, therefore avoiding to introduce any related assumption.
357 However, the statistical properties of the model error still need to be deciphered, although
358 in a less detailed (and perhaps non-parametrical) manner, to compute the integrals in eq.
359 (8) and (9) (see Section 4 for details on application). In the applications presented in
360 this paper we use the meta-Gaussian approach by *Montanari and Brath* [2004] to this
361 end. Its robustness notwithstanding, further research studies are needed to provide an
362 efficient statistical characterisation of the errors from hydrological models, which are often
363 heteroscedastic and affected by several forms of dependence not easy to decipher (see, for
364 instance, *Refsgaard et al.* [2006], *Kavetski et al.* [2006] and *Beven* [2006]). The interested
365 reader is also referred to *Montanari and Grossi* [2008] for an additional discussion on the
366 meta-Gaussian approach and error dependency. It is important to note that the model
367 error should be representative of model performances in validation.

4. Application of the Proposed Framework: Integrating Hydrological Model Implementation and Uncertainty Assessment

368 Estimating the probability distribution of the true value of the variable to be predicted
369 by a hydrological model (prediction uncertainty) is equivalent to simultaneously carrying
370 out model implementation and uncertainty assessment. The framework for estimating the
371 probability density function of model prediction, $f_Q(Q)$, was described in Section 3. Here
372 we show how eq. (8) can be applied in practice.

373 We assume that the probability density functions of model parameters, input data and
374 model error are known, for instance because they were already estimated by using proce-
375 dures such as those found in the hydrological literature (see, for instance, *Clark and Slater*
376 [2006], *McMillan et al.* [2011], *Renard et al.* [2011] and *Di Baldassarre and Montanari*
377 [2009] for data uncertainty; *Vrugt et al.* [2007], *Ebtehaj et al.* [2010] and *Srikanthan et al.*
378 [2009] for parameter uncertainty; and *Montanari and Brath* [2004], *Montanari and Grossi*
379 [2008] and *Krzysztofowicz* [2002] for model error). A practical demonstration showing
380 how this can be determined is contained in Section 6 below. Some of the above mentioned
381 techniques require the estimation of model likelihood, which may imply approximations in
382 the definition of the related uncertainties. For instance, likelihood assessment is required
383 by DREAM, which is used in the applications presented in Section 6 to estimate param-
384 eter uncertainty. However, methods are available to avoid the use of the likelihood to
385 define the above densities. For instance, bootstrap and resampling methods can be used
386 for parameter uncertainty [*Ebtehaj et al.*, 2010; *Srikanthan et al.*, 2009]. It is important to
387 note that definition of the above densities is a key task as an imperfect estimation of input
388 and parameter uncertainty would propagate through the simulation chain thus inducing
389 lack of fit. It is well known that the identification of such distributions is still a challenging
390 task in hydrology. In particular, for the data the problem still remains a relevant research
391 issue. Information on observation error, and the related probability distribution, can be
392 used to this end (see, for instance, *Pelletier* [1987]).

393 Under the above premises the double integral in eq. (8) can be easily computed through
394 a Monte Carlo simulation procedure, which can be carried out in practice by performing
395 many implementations of the deterministic hydrological model $S(\Theta, X)$.

396 In detail the simulation procedure is carried out through the following steps that refer
397 to the flowchart in Figure 1:

398 1. A parameter vector for the hydrological model is picked up at random from the
399 model parameter space according to the probability density $f_{\Theta}(\Theta)$.

400 2. An input data vector for the hydrological model is picked up at random from the
401 input data space according to the probability density $f_X(X)$.

402 3. The hydrological model is run and a model prediction (or a vector of predictions)
403 $S(\Theta, X)$ is computed.

404 4. An outcome of the model error (or vectors of errors) is picked up at random from
405 the model error population according to the probability density $f_e(e)$ and added to the
406 model prediction $S(\Theta, X)$.

407 5. The simulation described by items from 1 to 4 is repeated j times. Therefore we
408 obtain j (vectors of) outcomes of the prediction Q .

409 6. Finally the probability density $f_Q(Q)$ is inferred through the realizations mentioned
410 in item 5.

411 It is important to note that j needs to be sufficiently large, in order to accurately
412 estimate the probability density $f_Q(Q)$. Generally, a good compromise of accuracy and
413 computational efficiency to find an optimal j value should be evaluated case by case.

414 Once the probability distribution of the true value to be predicted Q is known we
415 obtain a best estimate for the prediction along with an assessment of uncertainty for a
416 given confidence level, under the above assumptions that are further discussed in the next
417 Section of the paper.

418 For the practical application of the multi-model approach, the whole simulation pro-
419 cedure is to be repeated for any subsequent candidate model therefore obtaining the
420 individual estimates for $f_Q^{(i)}(Q)$ to be averaged according to eq. (10).

421 In principle, the above framework allows one to estimate the single contribution of each
422 uncertainty source. For instance, if we are interested in the impact of parameter uncer-
423 tainty, the simulation procedure can be carried out by skipping items 2 and 4, therefore
424 neglecting the impact of data uncertainty and model error. However, we should be fully
425 aware that neither in the proposed framework nor in the real world are uncertainties nec-
426 essarily additive. Thus, even if assessment of individual impacts is possible, these latter
427 cannot be summed up a posteriori to assess the overall prediction uncertainty.

428 The algorithm presented above has some similarity with the operational flow chart of
429 other simulation methods like GLUE or multi-model approaches. The relevant difference
430 is the use of the model error to summarize uncertainties other than those induced by
431 imprecise input data and parameters. In this way likelihood computation can be avoided.
432 We stress once again that, in order to preserve the statistical consistency that is ensured
433 by the underlying theoretical development, the probability distribution of the model error
434 must be reliably inferred with the support of statistical tests.

5. Discussion of the Underlying Assumptions

435 Like any scientific method, the blueprint proposed in Sections 3 and 4 is based on
436 assumptions in order to ensure applicability. When dealing with uncertainty assessment
437 in hydrology, assumptions are often treated with suspicion, because it is felt that they
438 undermine the effectiveness of the method and therefore its efficiency and credibility with
439 respect to users. We stress here that assumptions (typically simplifying ones) are a means

440 to reach a better understanding of the behaviors of natural processes and allow science
441 to be effectively put into practice. As a matter of fact, assumptions are unavoidably
442 needed to set up models, calibrate their parameters and estimate their reliability, whatever
443 approach is used. Evidently, flawed assumptions may falsify statistical inference as well
444 as any alternative model of uncertain and deterministic systems. Therefore the target of
445 the researcher should not be to avoid assumptions, but rather discuss them transparently,
446 evaluate their effects and, when possible, check them, for instance through statistical
447 testing.

448 In order to discuss the assumptions conditioning the blueprint we introduced above,
449 first we note that we assumed that the uncertainty of model outputs only depends on
450 input data uncertainty, parameter uncertainty and model error, according to eq. (1)
451 which states that model output itself only depends on input data, parameters and model
452 structure S . Therefore, some sources of uncertainty are not taken explicitly into account,
453 like for instance operation uncertainty [*Montanari et al.*, 2009] and discretisation errors
454 when dealing with daily data. Indeed, several uncertainties are not explicitly accounted for
455 in any uncertainty assessment method. To this regard, we would like to point out that the
456 model error, for a given model and given input data and parameters, implicitly takes into
457 account, in an aggregated and very practical form, all the sources of uncertainty that make
458 the model output different with respect to observed values. However, other uncertainties
459 sources, like for instance uncertainty in the initial conditions and state variables, could
460 be included explicitly, provided the related probability distributions are quantified and
461 random outcomes for their values are randomly picked up at each simulation step. We
462 did not include additional uncertainties to simplify notation.

463 Second, we assumed that parameter uncertainty is independent of input uncertainty. If
464 the data are sufficient, this assumption is reasonable, because parameters of a given model
465 depend on statistics of the input data and not their particular values [*Casella and Berger,*
466 1994]. A practical demonstration of the limited sensitivity of rainfall-runoff model output
467 to artificially induced input errors was recently given by *Montanari and Di Baldassarre*
468 [2012]. We must note, however, that the data are seldom of sufficient size when fitting
469 hydrological models. As a result, parameters often turn out to be dependent on the input
470 uncertainty (so that changes in input data result in changes of parameters).

471 Further assumptions may be needed to estimate the probability density f_e of model er-
472 ror, which might be non-Gaussian and affected by heteroscedasticity. For instance, in the
473 application presented in Section 6 the meta-Gaussian approach by *Montanari and Brath*
474 [2004] is applied. In brief, the method recognizes the dependence on model prediction of
475 the conditional probability distribution of model error. In this way change of the statis-
476 tical properties during time is efficiently modeled and therefore the marginal probability
477 distribution of the model error is allowed to be heteroscedastic (see Section 6 and *Monta-*
478 *nari and Brath* [2004]). The meta-Gaussian approach assumes that the model error does
479 not depend on parameter and data uncertainty. In this case also, if the data are sufficient
480 the assumption is reasonable. We checked it with extended simulation experiments that
481 are independently presented by *Montanari and Di Baldassarre* [2012].

482 In principle the above assumptions of independence could be removed by conditioning
483 the model error and parameter uncertainty on data. Parameter uncertainty can be condi-
484 tioned by calibrating the hydrological model for different outcomes of the input data from
485 the related probability distribution. The model error can be conditioned by estimating

its probability distribution at each step of the simulation procedure described in Section 4, therefore obtaining different error probability densities for different input data and parameters. The main problem with this solution is given by the increased computational requirements. We further discuss this issue in Section 6.4.

If other approaches are used to derive the probability distribution of the model error, different assumptions would be introduced depending on the (possibly informal) approach that is adopted. No matter which method is used, any additional assumption introduced for inferring f_e should be appropriately checked.

Another relevant issue has been pointed out by some researchers (see, for instance, *Beven et al.* [2011]) who are convinced that epistemic errors arising from hydrological models might be not aleatory and therefore are difficult (or impossible) to model by using stochastic approaches. In our view, variables are either deterministic or random. That is, if they cannot be described deterministically, then they can be modeled by using stochastics, no matter if their stochastic dynamics are driven by epistemic uncertainty or natural variability. Another issue that is frequently raised is that epistemic errors are affected by non-stationarity and therefore cannot be efficiently modeled by using stochastics [*Beven and Westerberg*, 2011]. Actually, such a view neglects the fact that even the definition of stationarity and nonstationarity relies on the theory of stochastic processes [*Koutsoyiannis*, 2006, 2011] and thus dealing with it is necessarily an issue of applying stochastics. In our opinion, non-stationarity might be necessary to enrol when environmental changes are present, but it is not induced by epistemic uncertainty. Irrespective of its origin, non-stationarity can be efficiently dealt with by using non-stationary stochastic processes [*Brockwell and Davis*, 1987], and by introducing and checking suitable assumptions. For

instance, the stochastic kernel introduced in eq. (3) and (4) is conditioned on the input data and therefore its marginal statistical properties are changing in time. In addition, in the case studies we present here the model error is allowed to be heteroscedastic and correlated. Indeed, as we mentioned above, the meta-Gaussian approach provides a conditional probability distribution of the model error that is changing in time depending on the simulated river flow [*Montanari and Brath, 2004*].

It is important to note that the blueprint proposed here relies much on data. Although probability distributions of input data, model parameter and model error could be estimated according to expert knowledge, data analysis is a fundamental requirement for assessing uncertainty. Therefore, particular attention should be paid to data collection and checking, to avoid as much as possible the use of disinformative observations (for a detailed discussion on this issue the interested reader is referred to *Beven and Westerberg [2011]*).

One may be concerned by the computational requirements of the proposed framework, especially when dealing with complex modeling approaches. For instance, spatially distributed models might require sampling from several parameters and might involve significantly longer computational times. This issue is indeed a matter of concern for any numerical integration procedure. It needs to be carefully considered in view of the length of the simulation period and the minimal number of simulated data points that is required to reliably infer the probability distribution of the model output.

One reviewer of this paper asserted that, strictly speaking, none of the above assumptions is satisfied. We believe that if we accept such an assertion in its generality, we would convict all models and perhaps all scientific disciplines except pure mathematics,

532 because all sciences that describe Nature seek to provide approximations of reality. It is
533 well known that all models are wrong, and, likewise, all assumptions are never strictly
534 satisfied. The purpose in modeling is to produce approximations of reality, which are
535 tested whether or not they are satisfactory. If they are not, then the model should be
536 changed, trying different model structures or perhaps relaxing some assumptions.

537 However, our practical experience suggests that the assumptions we introduced are rea-
538 sonable. For instance, we believe that data uncertainty is indeed playing a negligible effect
539 on parameter uncertainty in most real world applications (see, for instance, *Montanari*
540 *and Di Baldassarre* [2012]).

6. Examples of Application

541 In order to illustrate the proposed blueprint with practical examples, two applications
542 are presented here below that refer to different rainfall-runoff models applied to catchments
543 located in Italy. In detail, the catchments are those of the Secchia River at Bacchello
544 Bridge and the Leo River at Fanano, in the Emilia-Romagna region, in Northern Italy.
545 Figure 2 shows their locations.

6.1. The case study of the Secchia River

546 The Secchia River is located in northern Italy and is a tributary to the Po River. The
547 catchment area is 1214 km² at the Bacchello Bridge river cross section that is located
548 about 62 km upstream of the confluence in the Po River. The maximum altitude is 2121
549 m above sea level (a.s.l.) at Mount Cusna. The main stream length up to Bacchello
550 Bridge is about 98 km and the climate over the region is continental.

551 Hourly rainfall and temperature data are available for the years 1972 and 1973 in five
552 raingauges. For the same period, hourly river flow data at Bacchello Bridge were collected.

553 To test the blueprint proposed here over an extended data set with controlled uncer-
554 tainty, we used synthetic hourly rainfall, temperature and river flow data that cover a
555 50-year observation period. The same data set was used by *Montanari* [2005] who gives
556 additional details. Synthetic data simulation is briefly described here below.

557 Rainfall data, for the 5 raingauges mentioned above, were generated using the gener-
558 alized multivariate Neyman-Scott rectangular pulses model [*Cowpertwait*, 1995] that was
559 calibrated using the observed data. Mean areal rainfall was then computed as a weighted
560 sum of the rainfall in each raingauge, where weights were estimated by using the Thiessen
561 polygons. Rainfall uncertainty was introduced through weight corruption by randomly
562 picking up their value, at each time step, from a uniform distribution in the range \pm
563 20% of the related uncorrupted value. The obtained weights were rescaled so that their
564 cumulative sum is equal to one.

565 Synthetic hourly temperature data were generated by applying a fractionally differenced
566 ARIMA model (FARIMA; see *Montanari et al.* [1997]). A mean areal value for temper-
567 ature was obtained by rescaling the synthetic observations to the mean altitude of the
568 basin area, by adopting a standard temperature gradient. Temperature data were not
569 corrupted, in view of their limited uncertainty with respect to rainfall and river flow.

570 Synthetic river flow data were generated by using the previously generated synthetic
571 rainfall and temperature records as input to the lumped rainfall-runoff model ADM [*Fran-*
572 *chini*, 1996]. The ADM model is a nine-parameter lumped conceptual scheme that was
573 calibrated against historical data obtaining a Nash efficiency [*Nash and Sutcliffe*, 1970] of

574 0.81 in validation (see *Montanari* [2005] for more details). Table 1 presents the model pa-
575 rameters. River flow data were corrupted by multiplying each observation by a coefficient
576 that was picked up, at each time step, from a uniform distribution in the range 0.8–1.2.
577 The coefficient of determination of the linear regression of corrupted versus uncorrupted
578 river flow data is 0.86.

579 The observations included in the first 30 years of the synthetic record were used to cali-
580 brate a rainfall-runoff model that has reduced complexity with respect to ADM, therefore
581 introducing model structural uncertainty (see below for a detailed description). Years from
582 31 to 40 were used to validate the model itself and to infer the probability distribution
583 of the model error by using the meta-Gaussian approach by *Montanari and Brath* [2004].
584 The goodness-of-fit was checked by using the statistical tests described in *Montanari and*
585 *Brath* [2004], which were satisfied over the whole range of river flows.

586 Finally, data for the years 41-50 of the observation period were used to test, in full
587 validation mode, the proposed blueprint (rainfall-runoff modeling and uncertainty assess-
588 ment).

589 The rainfall-runoff model we used for the Secchia River is HyMod, namely, the same
590 5-parameter lumped and conceptual rainfall-runoff model that was used by *Montanari*
591 [2005]. HyMod was introduced by *Boyle* [2000] and extensively used thereafter. Model
592 parameters are shown in Table 1. Evapotranspiration is accounted for by using the ra-
593 diation method [*Doorembos et al.*, 1984]. With a total of only five parameters, HyMod
594 can be considered an approach of reduced complexity with respect to ADM and therefore
595 model structural uncertainty is introduced.

HyMod was calibrated by using DREAM [Vrugt *et al.*, 2007], in which a standard Gaussian likelihood function was used. DREAM is a modified SCEM-UA global optimisation algorithm [Vrugt *et al.*, 2003]. It makes use of population evolution like a genetic algorithm together with a selection rule to assess whether a candidate parameter set is to be retained. The sample of retained sets after convergence can be used to infer the probability distribution of model parameters. Herein, a number of 6000 parameter sets were retained, which indirectly determine the density function $f_{\Theta}(\Theta)$ of the parameter vector in a non-parametric empirical manner, fully respecting the dependencies between different parameters. HyMod explained about 81% and 82% of the river flow variance in calibration and validation, respectively, with the best DREAM parameter combination from the joint Markov chains. The values of the corresponding Nash efficiency are 0.81 - 0.82. These are feasible values in real world applications. Figure 3 reports a comparison over a 1500 hour window of the full validation period (years 41-50) between observed and simulated hydrographs.

6.2. The simulation procedure for the Secchia River

As we mentioned above, the simulation procedure refers to the years 41-50 of the observation period. HyMod was run 5000 times, by randomly picking up parameter sets from those retained by DREAM and accounting for input uncertainty by corrupting, for each simulation, the rainfall data as described in Section 6.1 (that is, by reproducing the same type of error that was introduced in the synthetic data set. Namely, a perfect uncertainty assessment for the rainfall data was assumed). A random outcome from the probability distribution of the model error was added to each observation, therefore obtaining,

617 for each simulated river flow, a sample of 5000 points that allows to infer the related
618 probability distribution.

619 Figure 4 shows the 95% prediction limits for the same 1500 hour window of the full
620 validation period mentioned above, along with the corresponding observations. By looking
621 at the overall prediction, one notes that 5.4% and 4.3% of the observations are located
622 above the upper and below the lower limit, respectively, against theoretical values of
623 2.5% (at the 95% confidence level). These results indicate a slight underestimation of the
624 band widths. Figure 5 shows a coverage probability plot (CPP), which gives information
625 on the accuracy of the uncertainty estimation. A placement of the points along the 1:1
626 line is expected. For more details on drawing and interpreting the CPP plot (which is
627 sometimes referred to as probability plot or Q-Q plot) see *Laio and Tamea* [2007]. For
628 the present case, we note that Figure 5 confirms the underestimation of the band width,
629 which nevertheless is scarcely significant in practice.

6.3. The case study of the Leo River

630 The catchment area of the Leo River basin at the closure section of Fanano is 64.4 km²
631 and the main stream length is about 10 km. The maximum elevation in the catchment is
632 the Mount Cimone (2165 m a.s.l.), which is the highest peak in the northern part of the
633 Apennine Mountains. The climate is continental.

634 Daily river flow data at Fanano are available for the period January 1st, 2003 - October
635 26th, 2008, for a total of 2126 observations. For the same period, daily mean areal
636 rainfall and temperature data over the catchment are available, as estimated by the Italian
637 National Hydrographic Service based on observations collected in nearby stations.

638 The observations collected from January 1st 2003 to December 31st 2006 were used
639 for calibrating the rainfall-runoff model, while the period January 1st 2007 - October
640 26th 2008 was reserved for its validation. We estimated the probability distribution of the
641 model error by referring to the first year of the validation period (2007), in order to obtain
642 an assessment of $f_e(e|\Theta, X)$ in a real world application. Finally, the period January 1st
643 2008 - October 26th 2008 was reserved for testing, in full validation mode, of the proposed
644 blueprint (rainfall-runoff modeling and uncertainty assessment).

645 The rainfall-runoff model is AFFDEF [Moretti and Montanari, 2007], a spatially-
646 distributed grid-based approach where hydrological processes are described with
647 physically-based and conceptual equations. AFFDEF counts 8 calibrated parameters,
648 which are described in Table 1. In order to limit the computational requirements, and in
649 view of the limited catchment area, the Leo river basin was described by using only one
650 grid cell, therefore applying a lumped representation. AFFDEF was calibrated by using
651 DREAM [Vrugt et al., 2007], again by using a standard Gaussian likelihood function.
652 Herein, a number of 32000 parameter sets was retained. Figure 6 shows the probability
653 density function for the model parameters. They all appear to be unimodal and well
654 defined. Parameter values are in agreement with what one would expect from AFFDEF
655 applications to similar catchments [Moretti and Montanari, 2007].

656 AFFDEF explained about 57% and 47% of the river flow variance in calibration and
657 validation, respectively. The values of the corresponding Nash efficiency are 0.59 - 0.36.
658 Figure 7 depicts a comparison during the validation period (2007 and 2008) between ob-
659 served and simulated hydrographs. This latter was obtained by using the best DREAM
660 parameter set. We can see that a significant uncertainty affects the model performance,

661 which is unlikely merely due to lumping the model at catchment scale. We are interested
662 in checking whether the proposed blueprint provides a consistent assessment of such un-
663 certainty. The probability distribution of the model error was again inferred by using the
664 meta-Gaussian approach, which provided a satisfactory fit for river flows greater than 0.5
665 m^3/s .

6.4. The simulation procedure for the Leo River

666 No information is available about input uncertainty. This is an important limitation
667 in many practical applications. In particular, input uncertainty is usually dominant in
668 real time flash-flood forecasting, where input rainfall to a rainfall-runoff model is usually
669 predicted to increase the lead time of the river flow forecasting. If a probabilistic prediction
670 for rainfall is available then input uncertainty can be efficiently taken into account in
671 the blueprint proposed above. Alternatively, input uncertainty can be estimated using
672 expert knowledge, Bayesian procedures like BATEA [*Kavetski et al.*, 2006] or conditional
673 simulation methods [*Clark and Slater*, 2006; *Götzinger and Bárdossy*, 2008; *Renard et al.*,
674 2011]. For the present application, in absence of any information and similarly to *Vrugt et*
675 *al.* [2008] and *Renard et al.* [2010], we introduced in each simulation and each data point
676 a rainfall multiplier that was picked up from a Gaussian distribution with unit mean and
677 standard deviation equal to 0.1.

678 A number $j = 5000$ of AFFDEF simulations were run during the 300-day full validation
679 period January 1st 2008 - October 26, 2008. A random outcome from the probability
680 distribution of the model error was added to each observation, therefore obtaining, for
681 each simulated river flow, a sample of 5000 points.

682 Figure 8 shows the 95% prediction limits for the full validation period, along with the
683 corresponding observations. The results confirm the relevant uncertainty that is antici-
684 pated by model performance. In fact, the prediction bands cover a large range of river
685 flows. The observations located above the upper and below the lower limit are 7.3% and
686 9.1%, respectively, for river flows greater than $0.5 \text{ m}^3/\text{s}$. In this case also, the width of
687 the prediction limits appears to be slightly underestimated. Figure 9 shows the CPP plot
688 for the 300-day full validation period, for river flows greater than $0.5 \text{ m}^3/\text{s}$. The slight
689 underestimation of the band width is confirmed.

690 It is interesting to inspect the reasons for the underestimation of the prediction limits
691 width. In fact, one may note in Figure 8 that the lower prediction limit is satisfactorily
692 estimated, with the only exception of the final part of the validation period where ob-
693 servations systematically fall slightly outside the limit itself. On the contrary, the upper
694 limit seems to be too large and too narrow for low and high flows, respectively. Further
695 information can be gained by comparing the probability density functions of observed
696 and simulated data. Given that such distribution, for the simulated data, depends on the
697 magnitude of the model prediction according to the assumptions of the meta-Gaussian
698 approach (see Section 5), the above comparison must be carried out by focusing on a
699 restricted range of river flows, for which distribution changes are negligible. Figure 10
700 reports the result of the comparison for the validation period and the low flow range be-
701 tween $2 \text{ m}^3/\text{s}$ (the observed mean) and $5 \text{ m}^3/\text{s}$. It can be seen that the overestimation
702 of the upper limit is confirmed. To further inspect this issue, the comparison was also
703 performed between the probability density functions of actual and simulated model error
704 for the validation period and the same low flow range. Results are shown in Figure 11.

705 One can see that the variability of the model error is overestimated as well. Therefore, it
706 appears that the unsatisfactorily assessment of the upper prediction limit for low flows is
707 mainly due to inefficient representation of the statistical properties of the model error by
708 the meta-Gaussian approach, which is induced by the limited extension of the calibration
709 period (the year 2007 only) that makes statistical analysis and testing scarcely efficient. In
710 practice, the data base is not extended enough for the method to recognize the variability
711 of the band width depending on the river flow magnitude. Then, the method tends to
712 predict constant band width thus resulting in overestimation and underestimation for low
713 and high flows, respectively.

714 Other reasons for the lack of fit could be improper characterisation of input and param-
715 eter uncertainty as well as failure of the fulfilment of the underlying assumptions and in
716 particular that of independence between the model error and parameter/data uncertainty
717 that is adopted by the meta-Gaussian approach. In fact, the statistical properties of the
718 model error were estimated by referring to the simulation obtained with the best DREAM
719 parameter combination from the joint Markov chains. Actually, the error behaviors in-
720 ferred in this way are not fully representative of suboptimal input data and parameter
721 vectors that are picked up randomly in the simulation procedure which may induce larger
722 errors for the given data set. To avoid this problem, two solutions can be used: (a) to
723 estimate the model error by referring to a parameter and data set that provides average
724 performances instead of the best ones. This approach is computationally efficient and
725 therefore preferable when computing resources are a matter of concern. (b) To infer the
726 statistical properties of the model error at each simulation step, therefore significantly
727 increasing computational requirements.

728 We believe that performances like those we obtained in the two case studies above
729 are sufficiently accurate for real world decision making in view of the consistency of the
730 related data base. For other cases the most appropriate solution should be decided after
731 considering the related practical needs.

7. Process-based stochastic modeling as a learning process

732 Uncertainty estimation allows one to quantitatively assess model reliability. If the
733 model is process-based, the correctness of the underlying schematizations can be effec-
734 tively checked by looking at the obtained prediction limits. In fact, these latter provide a
735 comprehensive picture of the probability distribution of the prediction error, for a given
736 confidence level and different river regimes. Therefore, prediction limits are a possible
737 mean to check the correctness of our understanding of the hydrological processes at a
738 given place. A closer look at the full probability distribution of the model error, for dif-
739 ferent flow regimes (an example that refers to low flows is presented in Figure 11) allows
740 one to complete the information about model failures in different hydrological situations,
741 therefore providing useful indications on possible adjustments at the model structure (for
742 a recent discussion on model structural adequacy see *Gupta et al.* [2012]). In particular,
743 the above distribution indicates when the model fails and the type of failure that is occur-
744 ring, so that its impact can be evaluated. The prediction error should be assessed by also
745 considering input and parameter uncertainty, to better understand whether the weakness
746 is related to model structure rather than calibration information. Analysis of parameter
747 uncertainty, as depicted by the parameter distributions shown in Figure 6, allows one to
748 assess whether the parameters themselves are well defined and what is their impact on
749 the results. A flat distribution may correspond to a poorly defined parameter and/or a

750 scarce impact of the related process on the results. Such analysis is particularly useful
751 when adopting a flexible model structure, to identify the relevant model components (for
752 applications, see *Montanari et al.* [2006]; *Fenicia et al.* [2008]; *Schoups et al.* [2008]).

8. Conclusions and discussion

753 A blueprint is presented to introduce a novel methodological scheme for estimating
754 the uncertainty of the output from a process-based (deterministic) hydrological model
755 through the estimation of the related probability distribution. The scheme is obtained
756 by developing a theoretical formulation to convert a deterministic hydrological model
757 into a stochastic one, therefore incorporating randomness in hydrological modeling as a
758 fundamental component. The scheme shows that to include an arbitrarily complex deter-
759 ministic model within a stochastic framework, where randomness is a fundamental part
760 of the system, is in principle possible. Although we explicitly focused on process-based
761 approaches, the blueprint that we are proposing is applicable to any deterministic scheme.
762 The relevant feature of the approach herein proposed, which can be applied to models of
763 arbitrary complexity, is that model likelihood computation can be avoided. In fact, the
764 approach proposed replaces the single output of a deterministic model with the proba-
765 bility distribution thereof which is estimated by stochastic simulation. A comprehensive
766 discussion of the underlying simplifying assumptions and how they can be removed is
767 presented, therefore allowing to structure modeling in a objective setting. The proposed
768 method allows hydrological modeling and uncertainty assessment to be jointly carried out.

769 Two applications are presented for illustrating the introduced blueprint. One of them
770 makes use of synthetic data. Although simplifying assumptions are introduced to reduce
771 the computational effort, the case studies show that the proposed approach is efficient

772 and provides a consistent uncertainty assessment. However, the results show that the
773 opportunity of removing some of the above assumptions should be considered depending
774 on the user needs.

775 We believe the theoretical framework introduced here may open new perspectives re-
776 garding modeling of uncertain hydrological systems. In fact, analyzing randomness within
777 process-based system representations is an invaluable opportunity to improve system un-
778 derstanding therefore increasing predictability, according to the “models of everywhere”
779 concept [*Beven, 2007*]. In particular, it is possible to analyse (possibly) changing or
780 shifting behaviors and their reaction to (human induced) changes. Moreover, we believe
781 that the proposed procedure is very useful for educational purposes, putting the basis for
782 developing a unified theoretical basis for uncertainty assessment in hydrology.

783 A successful application of the proposed blueprint requires a reliable estimation of input,
784 parameter and model uncertainty. The latter is obtained through the estimation of the
785 probability density $f_e(e)$ of the model error. The meta-Gaussian model [*Montanari and*
786 *Brath, 2004; Montanari and Grossi, 2008*] was herein used. However, in condition of
787 data scarcity it may be scarcely efficient, as we show in Section 6.4. Data assimilation
788 methods can also be considered, like machine learning and nearest neighbor techniques
789 [*Shrestha and Solomatine, 2006*]. All the above methods rely on limiting assumptions and
790 some of them are also computer intensive. We believe that estimating model uncertainty in
791 hydrology is still a difficult problem for which more focused research is needed [*Montanari,*
792 *2011*]. The proposed framework may facilitate streamlining of this research and linking it
793 with other components within an holistic modeling approach.

794 Finally, as mentioned above, estimation of parameter and input uncertainty is a relevant
795 challenge as well which has an impact on model prediction. Possibilities are the GLUE
796 method [*Beven and Binley, 1992*] and the DREAM algorithm [*Vrugt et al., 2007*] for
797 parameter uncertainty, which nevertheless are computer intensive as well and may turn
798 out to be impractical with spatially-distributed models applied to fine time scale at large
799 catchments. Information on observation error, and the related probability distribution,
800 can be used to estimate input uncertainty. Additional and focused research is needed
801 to improve the above techniques therefore ensuring a more practical application of the
802 framework herein proposed.

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Table 1. Parameters of the ADM, HyMod and AFFDEF rainfall-runoff models. Symbols are given for the parameters of AFFDEF which refer to Figure 6.

Parameter	Unit	ADM	HyMod	AFFDEF	Symbol
Maximum soil storage capacity	[cm]	X	X		
Shape parameter of the storage capacity curve	[-]	X	X		
Surface/subsurface flow partition factor	[-]		X		
Residence time quick flow reservoir	[h]		X		
Residence time low flow reservoir	[h]	X	X	X	K
Shape parameter of the drainage curve	[-]	X			
Shape parameter of the percolation curve	[-]	X			
Maximum drainage rate	[cm/s]	X			
Maximum percolation rate	[cm/s]	X			
Convectivity	[cm/s]	X			
Diffusivity	[cm ² /s]	X			
Multiplying factor for soil storativity	[-]			X	H
Multiplying factor for interception storage	[-]			X	C _{int}
Residence time soil water	[h]			X	H _s
Threshold temperature for snow accumulation	[°C]			X	T _s
Threshold temperature for snow melting	[°C]			X	T _{melt}
Snow conversion factor	[-]			X	SCF
Melting factor	[mm/(°C d)]			X	M _f

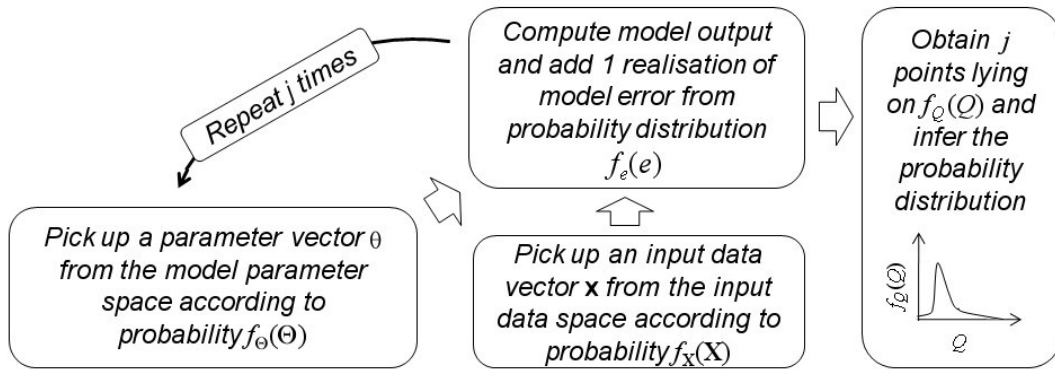


Figure 1. Flowchart of the Monte Carlo simulation procedure for performing the numerical integration in eq. (8) and (9).



Figure 2. Location of the case study basins (Italy). A and B indicate the positions of Bacchello Bridge and Fanano, respectively.

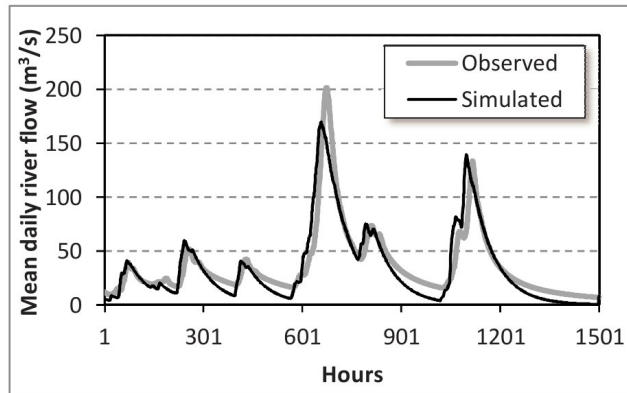


Figure 3. Case study of Secchia River. Comparison between observed and simulated hydrographs during a 1500-hour window included in the full validation period (years 41-50 of the synthetic record). The simulated hydrograph was obtained by using the best DREAM parameter set during the calibration period.

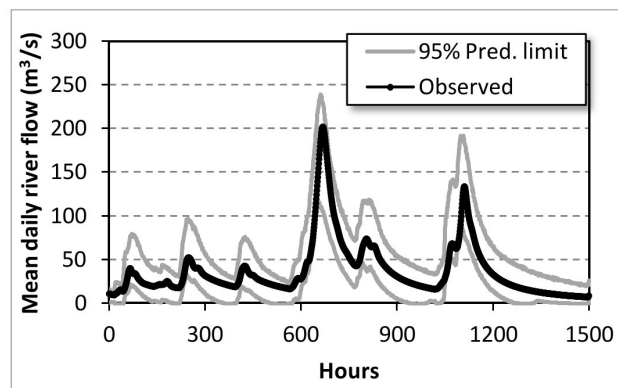


Figure 4. Case study of Secchia River. 95% prediction limits provided by HyMod during a 1500-hour window of the full validation period (years 41-50 of the synthetic record).

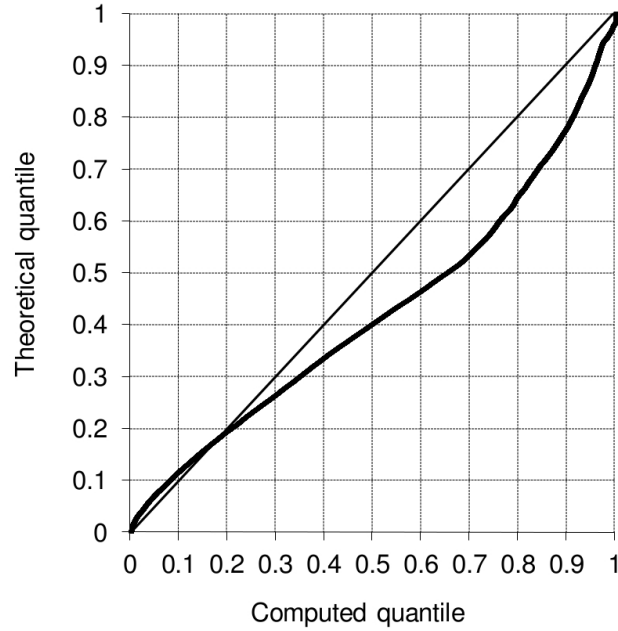


Figure 5. Case study of Secchia River. CPP plot of the prediction provided by HyMod during the full validation period (years 41-50 of the synthetic record).

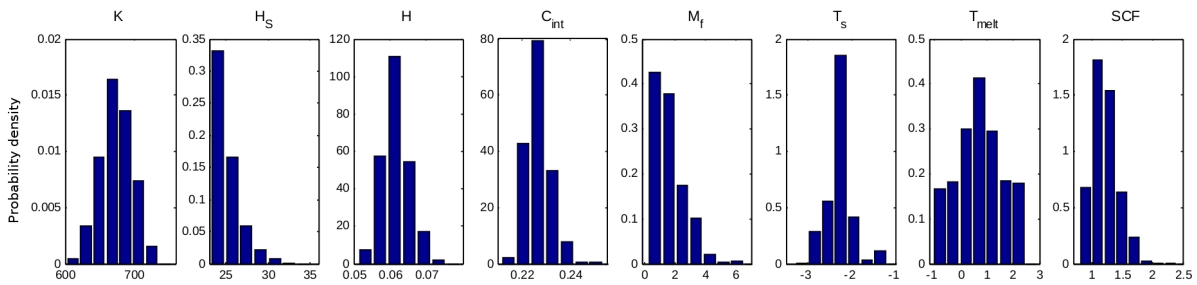


Figure 6. Case study of Leo River. Probability density functions for the AFFDEF parameters obtained with DREAM. Symbol meanings are given in Table 1.

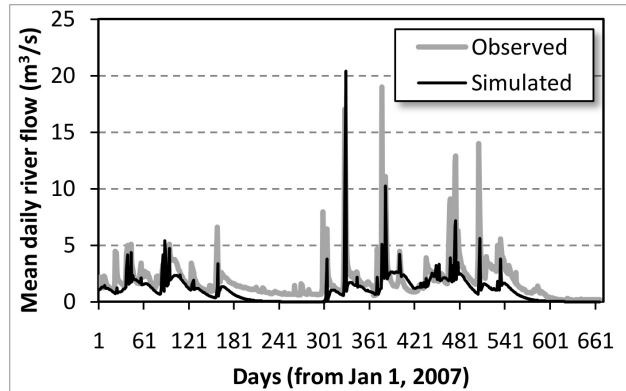


Figure 7. Case study of Leo River. Comparison between observed and simulated hydrographs during the validation period (Jan 1st 2007 - October 26, 2008). The simulated hydrograph was obtained by using the best DREAM parameter set during the calibration period.

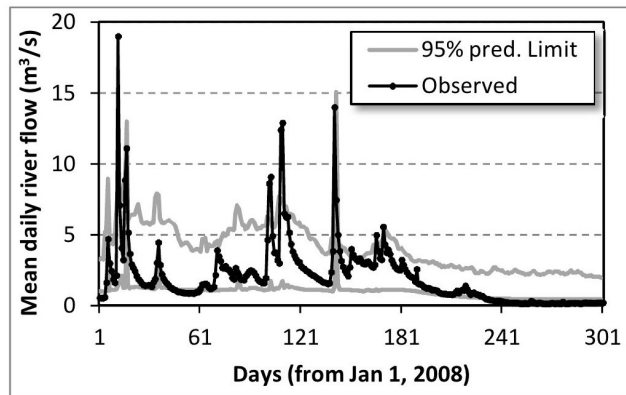


Figure 8. Case study of Leo River. 95% prediction limits provided by AFFDEF during the full validation period (Jan 1st, 2008 - Oct 26th, 2008), along with the corresponding observed values.

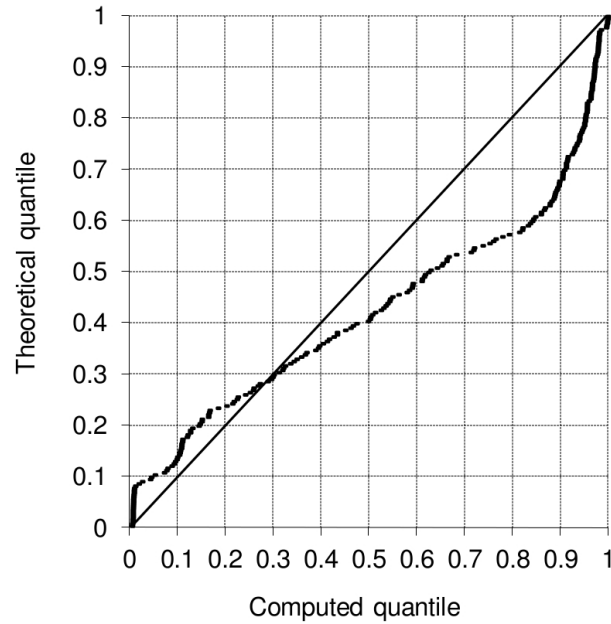


Figure 9. Case study of Leo River. CPP plot of the prediction provided by AFFDEF during the full validation period (Jan 1st, 2008, Oct 26th, 2008).

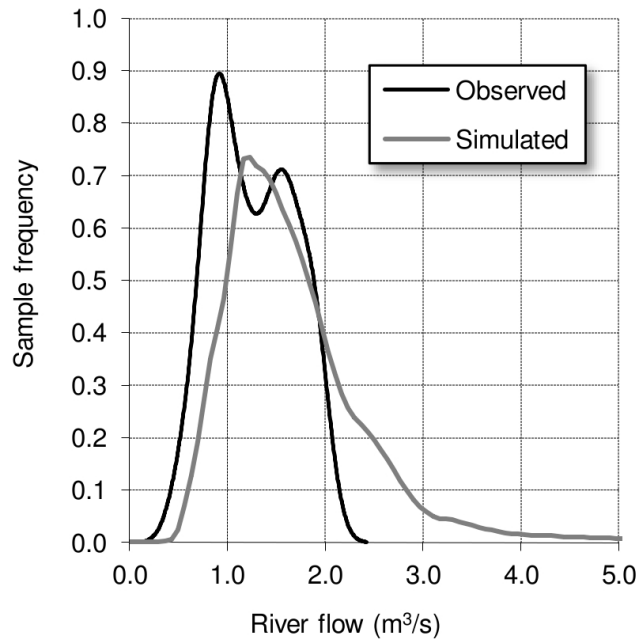


Figure 10. Case study of Leo River. Comparison between the probability density functions of observed and simulated data during the full validation period (Jan 1st, 2008, Oct 26th, 2008) for the river flow range between $2 \text{ m}^3/\text{s}$ and $5 \text{ m}^3/\text{s}$.

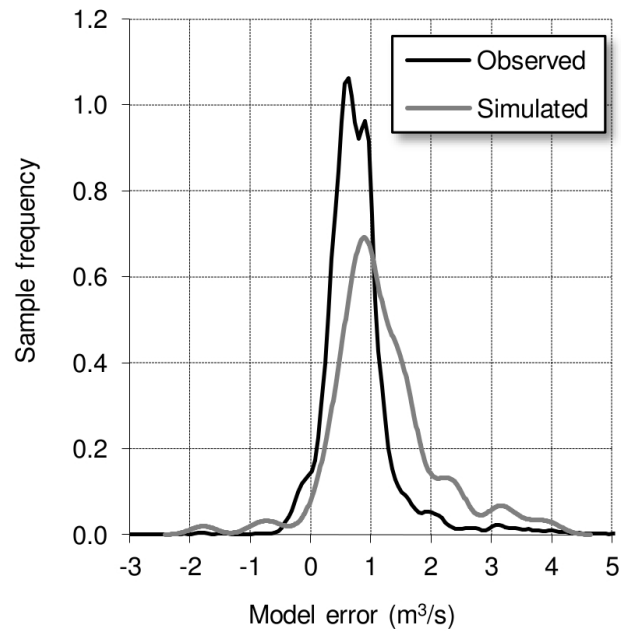


Figure 11. Case study of Leo River. Comparison between the probability density functions of actual and simulated model error during the full validation period (Jan 1st, 2008, Oct 26th, 2008) for the river flow range between 2 m³/s and 5 m³/s.